

1 **Title:** Continental-scale assessment reveals inadequate monitoring for threatened vertebrates in a  
2 megadiverse country

3

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25

26 **Abstract**

27 Monitoring threatened species is essential for quantifying population trends, understanding  
28 causes of species' declines, and guiding the development and assessment of effective recovery  
29 actions. Here, we provide a systematic, continental-scale evaluation of the extent and quality of  
30 monitoring for threatened species, focussing on terrestrial and freshwater vertebrates in  
31 Australia. We found marked inadequacies: one in four threatened taxa are not monitored at all;  
32 for taxa that are monitored, monitoring quality, as assessed across nine metrics, was generally  
33 low. Higher quality monitoring was associated with policy recognition, in the form of species  
34 recovery plans, and for species having a more imperilled conservation status. Across taxonomic  
35 classes, the proportion of species monitored was highest for mammals and then birds, whereas  
36 monitoring quality was greatest for birds. Improving monitoring quality requires setting clear  
37 objectives, direct integration with management, incorporating explicit management triggers,  
38 long-term resourcing, and better communication and accessibility of monitoring information.  
39 While our results revealed that overall monitoring efforts are inadequate, the positive  
40 relationship between improved monitoring outcomes and national policy support highlights that,  
41 when resources are available, good monitoring outcomes can be achieved. Quality monitoring  
42 programs for threatened species, and biodiversity more generally, should be recognized as vital  
43 measures of a nation's progress, analogous and complementary to more widely-used economic  
44 and human health indicators.

45

46 **Key words:** adaptive management; conservation; conservation policy; extinction, management;  
47 monitoring; threatened species

48

49

## 50 **Introduction**

51 Monitoring threatened species is crucial to halting biodiversity loss (Legge et al., 2018; Primack,  
52 2006). Information on species population trajectories is essential for assessing extinction risk,  
53 determining species' responses to threatening processes, prioritizing remedial management,  
54 evaluating management effectiveness (Balmford et al., 2005; Legge et al., 2018; Marsh and  
55 Trenham, 2008), and improving understanding and management of threats (Garnett et al., 2018).  
56 In contrast, absence of robust information on species' trajectories can lead to poor allocation of  
57 conservation resources (Campbell et al., 2002; Marsh and Trenham, 2008; Robinson et al.,  
58 2018), sub-optimal conservation outcomes, and potentially leads to preventable extinctions  
59 (Lindenmayer et al., 2013; Woinarski et al., 2017). Aggregation of adequate monitoring data and  
60 synthesis of trends across species is also a pivotal requirement for assessment of policy  
61 performance (Loh et al., 2005; Tittensor et al., 2014).

62         Despite recognition that monitoring is crucial to conservation, it is rarely prioritized in  
63 threatened species management, and may be absent or of poor quality (Field et al., 2007; Legg  
64 and Nagy, 2006; Lindenmayer and Likens, 2018). For example, while endangered species  
65 recovery plans in the United States commonly provide for monitoring of target species'  
66 population trajectories, few consider threat, demographic or habitat trends (Campbell et al.,  
67 2002). Likewise, many monitoring programs have limited power to detect changes in abundance  
68 (Marsh and Trenham, 2008), or are not linked to management actions, resulting in situations  
69 where species are monitored to extinction (Lindenmayer et al., 2013).

70         Frameworks and principles to guide development of effective biodiversity monitoring  
71 programs have been proposed (e.g. Lindenmayer and Likens, 2009; Reynolds et al., 2016;  
72 Robinson et al., 2018). Key recommendations include ensuring monitoring: (1) aims to answer

73 clearly defined questions; (2) has clearly stated objectives and links to policy and management;  
74 (3) is underpinned by rigorous statistical design; and (4) can evolve iteratively in response to  
75 new information, research questions and technology (Lindenmayer and Likens, 2009; Reynolds  
76 et al., 2016; Robinson et al., 2018). Here we provide the first continental-scale systematic  
77 evaluation of the extent to which current monitoring complies with these recommendations for  
78 all threatened terrestrial and freshwater vertebrates in Australia.

79 We focused on Australia for two reasons. First, Australia is a megadiverse continent-  
80 country, with a broad range of species and ecosystems. Second, Australia has a poor track-record  
81 of halting species decline and extinction (Woinarski et al., 2015; Woinarski et al., 2017). Given  
82 the key role of monitoring in threatened species management (Legge et al., 2018), an assessment  
83 of current monitoring against the attributes that characterise high quality monitoring is an  
84 important step towards reversing declines. Our analysis identifies key deficiencies in current  
85 monitoring efforts for threatened species and their consequences, differences in monitoring  
86 extent and quality across taxonomic groups, and factors that are associated with higher quality  
87 monitoring. Building on these insights, we provide recommendations for improving monitoring  
88 for threatened species.

89

## 90 **Material and methods**

### 91 *Framework to assess monitoring extent and quality*

92 We used an assessment framework to consistently score the extent and quality of monitoring  
93 programs for each threatened species (Woinarski 2018; Table 1). The framework comprised nine  
94 metrics, each scored on a 0 (no monitoring) to 5 (optimal monitoring) scale (see Tables S2-S10  
95 for scoring criteria). Monitoring was defined as targeted, repeated survey efforts. Where multiple

96 monitoring programs were identified for a taxon, the evaluation metrics were scored from a  
97 national perspective on the aggregated/combined monitoring effort, so each taxon received a  
98 single monitoring score for each metric. Our assessment of monitoring was undertaken from July  
99 2016 to July 2017.

100

### 101 *Collating information*

102 We assessed monitoring for all Australian threatened vertebrates, excluding marine fish  
103 and marine mammals, listed as Critically Endangered, Endangered, or Vulnerable under the  
104 Australian Government's *Environment Protection and Biodiversity Conservation Act 1999*  
105 (EPBCA). We also assessed monitoring for some taxa that are not currently EPBCA-listed, but  
106 are assessed as threatened under State/Territory legislation, or by the International Union for  
107 Nature Conservation (IUCN), or other non-statutory listings. We refer to the conservation status  
108 of these taxa as 'Other'. For example, for fish, 19 taxa were categorised as 'Other'; these taxa  
109 have been assessed by the Australian Society for Fish Biology as nationally threatened, using  
110 IUCN listing criteria. For information on the number of taxa from each taxonomic group in each  
111 EPBCA listing category, and the number of Other taxa assessed, see Table S1.

112 Species in each taxonomic class were assessed by one or more of the authors with  
113 expertise for that class, using published information, personal communications with individuals  
114 involved in management, and information from relevant government agencies and non-  
115 government conservation organizations. Our assessments of monitoring for each taxonomic  
116 group was largely based on information collated during recent reviews of the conservation status  
117 of Australian birds (Garnett et al., 2011), mammals (Woinarski et al., 2014a), reptiles (Chapple  
118 et al., in press), frogs (unpubl.) and fish (unpubl.). These reviews included inputs from all

119 relevant researchers, state agencies and conservation NGOs about population status and trends,  
120 and for older reviews, was updated for this paper. Information was based on monitoring  
121 programs, with the characteristics of these programs described by their practitioners. Where  
122 contributors to these accounts indicated that no trend information was available, we contacted all  
123 relevant experts to confirm the absence of monitoring programs, or for details of any monitoring  
124 programs that were present, but could not provide such trend information. Notwithstanding our  
125 efforts, some monitoring activity for some taxa may have been overlooked, as information on  
126 monitoring is often not published and is sometimes obscure, potentially resulting in  
127 underestimation of monitoring effort. However, we believe it is unlikely that any such missing  
128 information would substantially alter our analyses and conclusions. To ensure consistency in  
129 scoring across taxonomic groups, the assessors thoroughly discussed the assessment framework  
130 before commencing assessment to ensure consistent interpretation and implementation.

131 We also collated information for each taxon's EPBCA recovery plan status. In Australia,  
132 threatened species recovery plans (typically lasting five years) are developed to facilitate and  
133 coordinate the recovery and conservation of threatened taxa. They have legislative powers but  
134 are not automatically mandatory for listed threatened taxa (see Walsh et al., 2013 for an  
135 overview of recovery planning in Australia). Taxa were categorized into three groups: (1)  
136 'current recovery plan', (2) 'lapsed recovery plan' or (3) 'never had a recovery plan'.

137

### 138 *Statistical analysis*

139 First, we quantified the proportion of taxa that receive some form of monitoring. We then  
140 investigated whether presence or absence of monitoring was associated with taxonomic class  
141 (amphibian, bird, fish, mammal and reptile), conservation status (EPBCA listing: Critically

142 Endangered, Endangered or Vulnerable, or ‘Other’), or EPBCA recovery plan status (‘current’,  
143 ‘lapsed’, or ‘none’). Our outcome variable was binary (presence (scores 1-5) or absence of  
144 monitoring (score 0)). We employed Bayesian logistic regression with the main effects of  
145 taxonomic class, conservation status, and recovery plan status as potential predictor variables.  
146 We constructed a set of eight potential models, which were then compared using the Leave-One-  
147 Out-Cross-Validation Information Criteria (LOOIC) (Vehtari et al., 2017). The most  
148 parsimonious model within two LOOIC of the best fitting model was selected as the best model.  
149 We report 95% credible intervals for model estimates and differences between the various levels  
150 of the categorical predictor variables.

151         In the second phase of our analysis, we focused only on taxa that received some form of  
152 monitoring identified in the first stage of our analysis. We investigated which of the above  
153 mentioned predictor variables (class, conservation status, and recovery plan status) influenced  
154 monitoring scores assessed for each of the nine metrics, and the total score summed across the  
155 nine metrics. We modelled scores using Bayesian linear models assuming a Gaussian  
156 distribution, and considered the same set of eight potential models, which were compared using  
157 LOOIC.

158         All analyses were conducted using Bayesian regression models in Stan (brms) package  
159 (Bürkner, 2016) in R (R Development Core Team, 2017). We used default priors (improper flat  
160 prior over the real line) for the regression parameters and a half Student-t with 3 degrees of  
161 freedom for the residual standard deviation in the linear model and Cauchy distribution with  
162 location zero and scale five for the logistic regression model parameters to avoid potential issues  
163 with complete separation. For each model, we ran four Markov Chains for 2000 iterations after



164 discarding the burn-in of 1000 iterations. All chains showed good mixing, as measured by the  
165 Gelman and Rubin convergence diagnostic (Gelman and Rubin, 1992).

166

## 167 **Results**

168 We assessed monitoring for 408 threatened Australian vertebrates (excluding marine mammals  
169 and marine fish), representing ~ 5.5% of the total number of described species in these classes  
170 (~7358: Walsh et al., 2013). We found that 303 (74%) threatened taxa received some monitoring,  
171 with the remainder not monitored at all. The proportion of species monitored was highest among  
172 mammals (89%), then birds (76%), amphibians (75%), reptiles (62%) and fish (53%). For  
173 monitored taxa, the average summed score across the nine metrics was 29 out of 45, with the  
174 highest average score for birds (32), followed by amphibians (31), fish (27), reptiles (25), and  
175 mammals (25). The mean scores for each assessment metric are summarized in Fig. 1.

176

### 177 *Extent and quality of threatened species monitoring*

178 The best ranked model for presence/absence of monitoring contained all three predictor  
179 variables: taxonomic class, conservation status, and recovery plan status (Table S11, Fig. S1).  
180 The predicted probability of monitoring was highest for mammals, followed by birds, reptiles,  
181 amphibians and fish (Fig. 2a). A higher proportion of taxa with current or lapsed recovery plans  
182 were monitored than for taxa that had never had a recovery plan (Fig. 2b). Likewise, Critically  
183 Endangered and Endangered taxa were more likely to be monitored than Vulnerable or Other  
184 taxa (Fig. 2c).

185

186 For taxa that were monitored, the best ranked model for total monitoring score also  
187 contained the three predictor variables: taxonomic class, conservation status, and recovery plan  
188 status (Table S11, Fig. S2). Predicted mean monitoring score was highest for birds, followed by  
189 amphibians, fish, mammals and reptiles (Fig. 3a). Species with lapsed recovery plans had the  
190 highest predicted scores, followed by species with current recovery plans, while scores were  
191 lowest for species with no recovery plan (Fig. 3b). Critically Endangered species had the highest  
192 predicted scores, followed by Other, with Vulnerable taxa having the lowest predicted scores  
193 (Fig. 3c; Fig. S3-S11 for the model predictions for each of the nine metrics).

194

## 195 **Discussion**

196 We conducted the first continental-scale evaluation of monitoring for a diverse array of  
197 threatened taxa, to identify the strengths and weaknesses of current monitoring efforts, and thus  
198 guide key improvements that could be made to prevent species loss. Our assessment revealed  
199 inadequacies in both the extent and quality of threatened species monitoring in Australia. One in  
200 four threatened taxa receives no monitoring. Where monitoring does occur, its quality (as  
201 assessed across nine metrics) is generally poor, with a low overall average score (29, out of a  
202 maximum of 45).

203

### 204 *Key deficiencies and consequences*

205 That one quarter of threatened Australian taxa are not monitored is symptomatic of a  
206 broader ad-hoc approach to threatened species conservation in Australia (Scheele et al., 2018),  
207 and is consistent with inadequate environmental monitoring in Australia (Cresswell and Murphy,  
208 2016). Notably, although not specifically targeting threatened species, the Australian Long Term  
209 Ecological Research Network was decommissioned in 2018 (Lindenmayer, 2017), further

210 eroding Australia's capacity to accurately assess species trajectories. Without monitoring, we are  
211 unable to assess extinction risk robustly, identify causes of decline, evaluate management  
212 effectiveness, identify species/population trends or trajectories, identify research priorities, or  
213 fully engage stakeholders and the community (Legg and Nagy, 2006; Lintermans, 2013b; Marsh  
214 and Trenham, 2008). Given our results, it is unsurprising that efforts to halt species declines in  
215 Australia have met with idiosyncratic and limited success.

216         Where taxa were being monitored, average scores were relatively low across the nine  
217 assessment metrics. Although scores for each metric were highly variable, four stood out as  
218 having particularly low values: (1) Design quality, meaning that monitoring had limited  
219 statistical power to detect changes in species abundance or site occupancy; (2) Demographic  
220 parameters, meaning that causes of decline, and critical life stages, would be hard to discern; (3)  
221 Data availability, meaning that any information collected was typically not publicly available,  
222 and (4) Management linkage, meaning the monitoring was not integrated with, nor informing  
223 management (Fig. 1). These metrics are those most likely to be severely limited by resource  
224 availability, and/or lack of expertise. Poor quality monitoring fails to deliver detailed knowledge  
225 of threat impacts and how they vary across environmental space and over time; information that  
226 is essential in successful recovery programs (Scheele et al., 2017).

227         We found that there was little publicly available information about, or data from,  
228 monitoring programs funded using public monies. Notwithstanding commitments in Australia's  
229 national biodiversity strategy (Commonwealth of Australia, 2016), there is no integrated  
230 monitoring program for biodiversity or threatened species in Australia, and no central location  
231 for storing monitoring information, or making such information publicly accessible (Legge et al.,  
232 2018). Consequently, the public has limited awareness of the trajectories of Australian threatened

233 species (typically negative), and hence relatively little reason for engagement and concern. In  
234 stark contrast, monitoring information on the performance of other public programs such as  
235 education or health are increasingly made available to the public, and the absence of monitoring  
236 is viewed as evidence of poor program governance (Lindenmayer et al., 2012).

237         Our assessment also highlights that current EPBCA lists of threatened terrestrial  
238 vertebrates and freshwater fish under-represent the number of taxa requiring  
239 recovery/conservation action (e.g. of the 56 fish considered in this review, only 38 are EPBCA  
240 listed). No or minimal monitoring for many, potentially most, non-listed taxa represents a hidden  
241 threat to biodiversity conservation in Australia. For some taxa with immediate and severe threats  
242 (e.g. >10 unlisted small-bodied galaxiid and rainbowfishes threatened by alien invasive species),  
243 extinction is possible before taxa are listed (Moy et al., 2018; Raadik, 2014). In many other  
244 cases, insufficient data inhibits assessment of conservation status (Walsh et al., 2013; Woinarski  
245 et al., 2014b). To overcome these limitations and provide early warnings of emerging declines,  
246 we also must monitor non-listed taxa (Lindenmayer et al., 2012). In particular, monitoring is  
247 needed for data-deficient species that are likely to be impacted by current or emerging threats.  
248 Citizen science, new technologies, and improved statistical analyses may help meet the challenge  
249 of increasing monitoring coverage for both threatened and non-threatened species (Lahoz-  
250 Monfort and Tingley, 2018).

251

### 252 *Factors associated with better monitoring*

253         Despite the poor overall monitoring scores in our assessment, we found that some species  
254 (e.g. Tasmanian devil, Leadbeater's possum, western swamp tortoise, orange-bellied parrot, red-  
255 finned blue-eye, orange-bellied frog) had exemplary monitoring for almost all metrics in our

256 framework, demonstrating that good monitoring programs are achievable. National policy and  
257 legislative support was associated with better monitoring: taxa with EPBCA recovery plans  
258 (either current or lapsed) were more likely to be monitored, and that monitoring was likely to be  
259 of higher quality. Taxa with lapsed plans still scored highly for monitoring quality, suggesting an  
260 enduring legacy of recovery planning; or that earlier plans, which were better supported by  
261 Australian government funding (Walsh et al., 2013), incorporated more rigorous monitoring.  
262 Monitoring quality was also higher for species with more imperilled conservation status,  
263 indicating that management and monitoring effort has been focused on species at highest risk of  
264 extinction. The snapshot nature of our assessment means that it is not possible to tease apart  
265 cause and effect between policy support and monitoring. For example, more imperilled species  
266 may elicit better monitoring; or more imperilled species may be easier to monitor (e.g. range-  
267 restricted, fewer to count); or good monitoring programs that provide robust information on  
268 extinction risk may support prompt and accurate listings.

269

#### 270 *Variation across taxonomic classes and countries*

271 Mammals and birds are more likely to be monitored, and monitored well, than other  
272 taxonomic groups, especially fish; a similar pattern of monitoring bias exists in Europe  
273 (Schmeller et al., 2009). There are several possible explanations for taxonomic biases. First,  
274 conservation resources and research are unevenly distributed across classes, with biases towards  
275 mammals and birds (Lawler et al., 2006; Walsh et al., 2013). In particular, reptiles and fish are  
276 underrepresented in EPBCA threatened species listings, meaning their monitoring may be under-  
277 resourced (Walsh et al., 2013). Second, some taxonomic classes are easier to monitor than  
278 others. For example, many threatened amphibians (which scored higher, on average, than

279 mammals, fish and reptiles) have restricted distributions and form conspicuous breeding  
280 aggregations, making them easier to monitor. Third, the currency and comprehensiveness of  
281 EPBCA lists varies among classes; for example, one third of fish taxa assessed as threatened by  
282 the Australian Society for Fish Biology are not listed under national legislation, which might  
283 contribute to lack of monitoring in this class (Lintermans, 2013a). Fourth, taxonomic groups  
284 have varying levels of buy-in from the public; birds are especially amenable to monitoring by  
285 community groups and have well-established public involvement in and programs for monitoring  
286 (e.g. Birdlife Australia's Birdata program). Our assessment focused on vertebrates, which are  
287 given disproportionately high attention in conservation management (Walsh et al. 2013): the  
288 status of monitoring for threatened invertebrates is likely to be even more parlous.

289         Comparing monitoring efforts among countries is challenging because publicly available,  
290 synthesised information on monitoring is limited (Schmeller et al., 2009). Notwithstanding,  
291 monitoring efforts for threatened species in Australia fall short of those undertaken in some  
292 countries. For example, in the United Kingdom, *State of Nature* reporting provides publicly  
293 available information on the trajectory of thousands of species (Hayhow et al., 2016). Similarly,  
294 monitoring actions are mandatory in recovery plans for threatened species in the United States  
295 (Campbell et al., 2002). More broadly, a general pattern of inadequate biodiversity monitoring  
296 has been reported across the majority of regions worldwide (Balmford et al., 2005).

297

### 298 *Improving threatened species monitoring*

299         Broad deficiencies in threatened species monitoring in Australia highlight a critically  
300 important and urgent need for a more robust and integrated approach. Improving both the extent  
301 and quality of threatened species monitoring is a necessary first step in efforts to redress

302 Australia's poor conservation record. Globally, under-funding remains an inescapable  
303 conservation challenge (Waldron et al., 2017). This challenge is particularly acute in Australia,  
304 where environmental spending is disproportionately low, with Australia one of only four  
305 developed countries featuring in the top 40 underfunded countries for conservation spending  
306 (Waldron et al., 2013). Further, biodiversity conservation has experienced sharp reductions in  
307 funding over the past decade, receiving less than five cents for every \$100 of Australian  
308 government spending in 2018 (ACF, 2018). To achieve effective conservation outcomes,  
309 Australia must increase spending on biodiversity conservation (Scheele et al., 2018). As long as  
310 recovery plans are the critical mechanism for guiding species recovery, then all recovery plans  
311 should include quality-assured and funded monitoring, as legislated in the United States under  
312 the USA Endangered Species Act (Campbell et al., 2002). The value of investing in monitoring  
313 is clearly demonstrated by the positive association between good-quality monitoring and the  
314 level of understanding and management of threats for threatened species (Garnett et al., 2018).

315         At the scale of individual monitoring programs, there is much that can be done to  
316 increase monitoring extent and quality, despite limited resources. (1) Monitoring needs to be  
317 closely linked with management, with clear objectives, and explicit triggers for responsive  
318 management actions. (2) Specified monitoring objectives should guide the methodological  
319 design of fit-for-purpose monitoring programs (Robinson et al., 2018). (3) Monitoring must be  
320 recognised as a long-term activity with secure resourcing, rather than an occasional ad-hoc  
321 activity undertaken when surplus resources become available, or after it has become apparent  
322 that management actions have failed. This could be achieved by prioritizing and mandating an  
323 adequate monitoring program within any recovery plan or equivalent management document. (4)  
324 Monitoring should be a mechanism for communication and engagement with all stakeholders,

325 with responsible agencies recognising an obligation to provide, interpret and disseminate  
326 monitoring results to all stakeholders, including the broader public. (5) Adequate attention must  
327 be given to data management and metadata collection. (6) A national program to facilitate the  
328 storage, analysis, interpretation of, and public accessibility to, monitoring data, is urgently  
329 needed (Legge et al., 2018). (7) Information from monitoring programs for threatened species,  
330 and biodiversity more generally, should be recognized as a vital measure of a nation's progress,  
331 analogous and complementary to the more widely-used economic and health indicators.

332

### 333 **Acknowledgements**

334 This research was funded by the Australian Government's National Environmental Science  
335 Program through the Threatened Species Recovery Hub. We thank everyone who helped  
336 compile monitoring information. In particular, the Australian Society for Fish Biology  
337 Threatened Fish Committee and Wayne Robinson helped undertake the fish assessment. BirdLife  
338 Australia and its members, plus numerous experts, provided information on birds. Andrew  
339 Burbidge and many other experts contributed to the mammal assessment.

340

### 341 **Supplementary material**

342 Table S1. Information on the number of species included in the assessment.

343 Tables S2-S10. Scoring criteria for each of the nine metrics used to assess monitoring quality.

344 Table S11. Leave-One-Out-Cross-Validation Information Criteria for each of the eight models  
345 considered for each of the 11 response variables.

346 Figures S1-S11. Model predictions for: presence/absence of monitoring, total score, and each of  
347 the nine metrics.



348

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452 Australian vertebrate species. *Conserv. Biol.* 31, 13-23.

453

454

455 **Table**

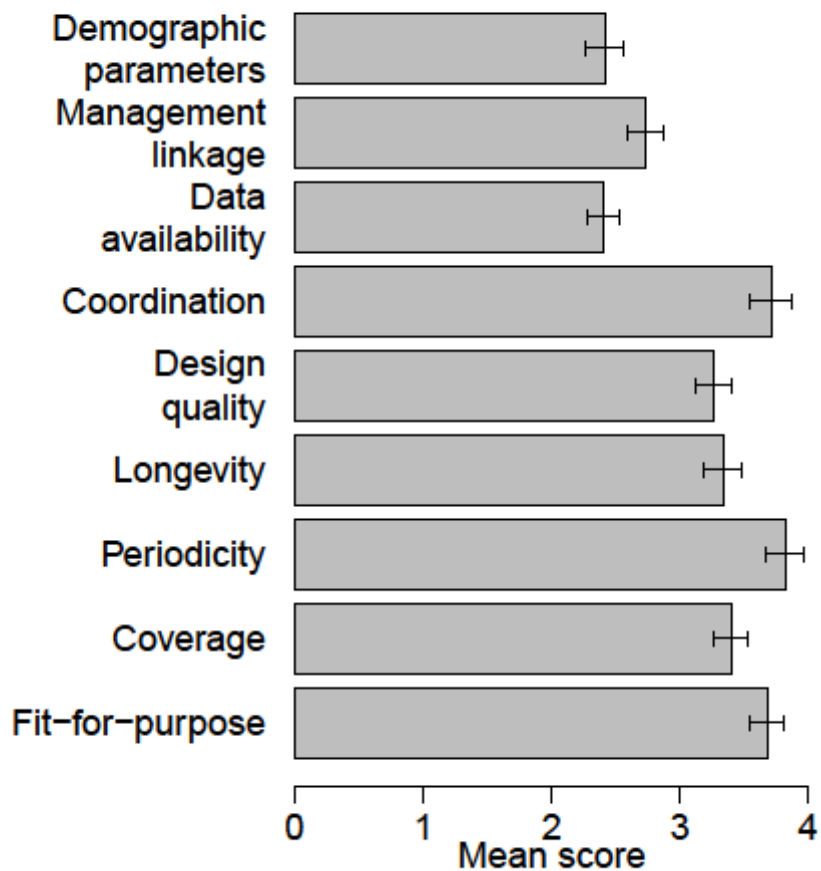
456 Table 1. Description and rationale for each of the nine metrics used to evaluate the quality of  
 457 threatened taxa monitoring adopted from Woinarski (2018). For each metric, taxa were scored 0-  
 458 5 (see scoring criteria, Tables S2-S10).

<b>Metric</b>	<b>Description</b>	<b>Rationale</b>
1. Fit-for-purpose	The use of methodologies designed to optimize detection of the target species.	To provide robust information, species-specific methods that consider the ecology and detectability of the target species are needed.
2. Coverage	The spatial extent of monitoring efforts across the target species' distribution.	A species' abundance and threat milieu can vary markedly across its distribution. As such, monitoring across a species distribution is needed to provide representative information on the species' trajectory.
3. Periodicity	Frequency of monitoring.	Timely information on a species' trajectory is needed. Monitoring should be undertaken frequently enough to be able to detect rapid changes and inform management.
4. Longevity	Longevity of monitoring.	Monitoring needs to be undertaken over sufficient timeframes to differentiate short-term variability from longer-term trends. Monitoring also needs to be able to identify small, incremental changes that may not be apparent where monitoring duration is limited.
5. Design quality	The statistical power of monitoring to detect trends in the occupancy/abundance of the target species.	Sufficient replication and detection frequency is needed to identify robust trends in the occupancy/abundance of the target species.
6. Coordination	The coordination of monitoring efforts among relevant jurisdictions and stakeholders.	When monitoring is performed by multiple organizations, its design, analysis and reporting needs to be effectively integrated to ensure comparable data are obtained.
7. Data availability and reporting	The availability and reporting of monitoring information.	For the value of monitoring data to be maximized, it must be readily accessible and well-curated, with adequate metadata and secure long-term storage.
8. Management linkage	Integration of monitoring and management actions.	Monitoring should inform the design and implementation of management, as well as be able to evaluate effectiveness.
9. Demographic parameters	The inclusion of demographic parameters in monitoring efforts.	In most cases, monitoring should involve assessment of critical demographic parameters, rather than just abundance. Information on life-history parameters can provide important ecological insights and help refine management.

459

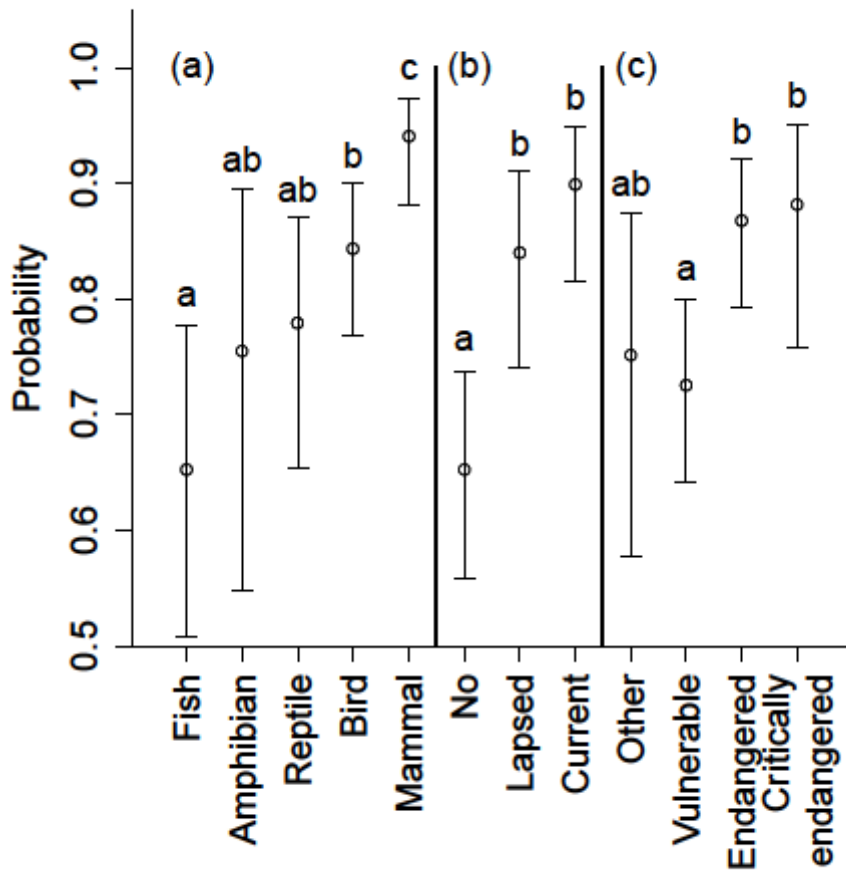
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461 **Figures**



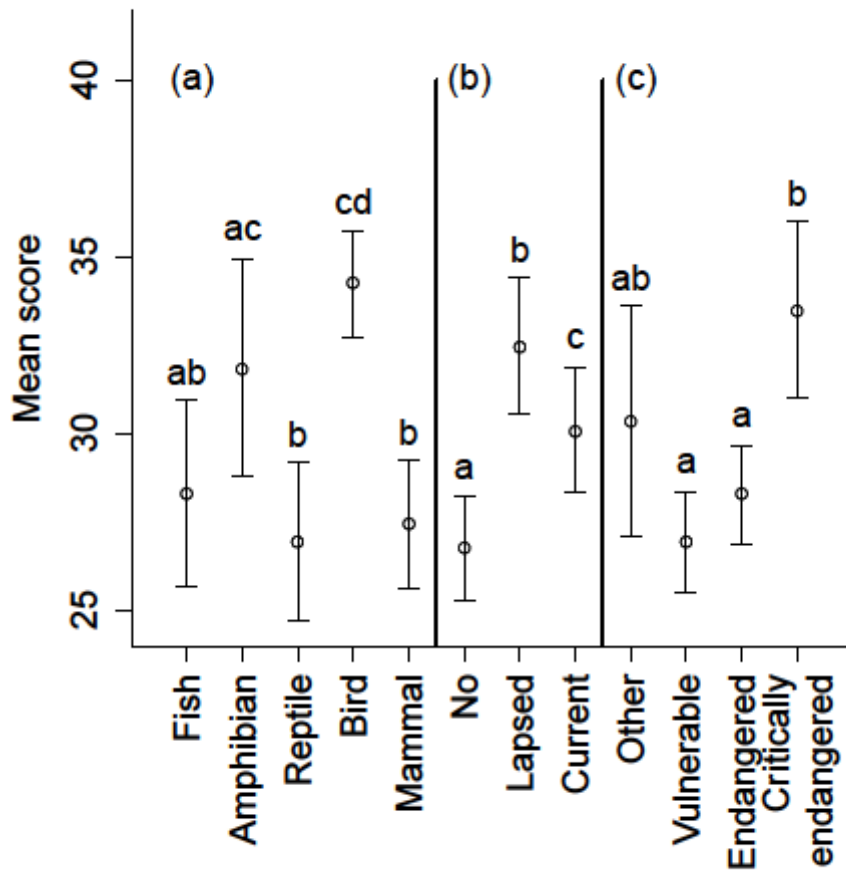
462  
463 Figure 1. Mean scores for each of the nine assessment metrics for monitored taxa. Error bars  
464 show the 95% credible intervals.

465



466  
 467 Figure 2. Probability of presence/absence of monitoring for Australian threatened taxa by (a)  
 468 taxonomic class, (b) recovery plan status, and (c) conservation status. In each case, the  
 469 probability of monitoring was predicted at average values for the other two predictors in the  
 470 model. Different letters (within a panel) indicate significant differences between predicted values  
 471 where the 95% credible interval for the log odds ratio does not cross zero.

472



473  
 474 Figure 3. Predicted mean total score for monitoring quality for Australian threatened taxa by (a)  
 475 taxonomic class, (b) recovery plan status, and (c) conservation status. In each case, predictions  
 476 were made at average values for the other two predictors in the model. Different letters (within a  
 477 panel) indicate significant differences between predicted values where the 95% credible interval  
 478 for the difference does not cross zero.

479



480 **Supplementary material**

481 **Article:** Continental-scale assessment reveals inadequate monitoring for threatened vertebrates  
482 in a megadiverse country

483  
484 **Table S1**, Information on the number of taxa in: (1) each taxonomic class; (2) each  
485 *Environmental Protection and Biodiversity Conservation Act 1999* (EPBCA) conservation status  
486 category; and (3) non-EPBCA listed taxa, that are considered threatened in non-statutory lists  
487 such as the International Union for the Conservation of Nature Red List of threatened species  
488 ('Other').

<b>Taxonomic class</b>	<b>Critically Endangered</b>	<b>Endangered</b>	<b>Vulnerable</b>	<b>Total</b>	<b>Other</b>	<b>Total</b>
<b>Fish (freshwater)</b>	5	15	17	37	19	56
<b>Amphibian</b>	5	14	10	29	0	29
<b>Reptile</b>	10	18	33	61	9	70
<b>Bird</b>	15	54	73	142	12	154
<b>Mammal</b>	5	33	57	95	4	99
<b>Total</b>	40	134	190	364	44	408

489

490 **Notes on sub-species, undescribed taxa and non-EPBCA listed species**

491 **Freshwater fishes:** Includes 19 taxa (including recognised but currently undescribed taxa) listed  
492 on the Australian Society for Fish Biology national list of threatened fish species, assessed in  
493 accordance with the International Union for the Conservation of Nature (IUCN) Red List criteria  
494 for threatened species listing <http://www.asfb.org.au/assets/ASFB/Threatened-Species-Committee/ASFB-2016-Threatened-Fishes-Committee-report.pdf>. Includes one sub-species  
495 currently listed by EPBCA. EPBCA listed marine species excluded.

497 **Amphibians:** Only *Environmental Protection and Biodiversity Conservation Act 1999* (EPBCA)  
498 listed species included. The conservation status of all Australian amphibians is currently being  
499 revised (as of June 2018).

500 **Reptiles:** Includes four species listed as threatened by the IUCN, but not included under the  
501 EPBCA. EPBCA listed marine species excluded.

502 **Birds:** Includes 12 species listed as threatened by the IUCN, but not included under the EPBCA.  
503 EPBCA total includes 11 sub-species that were assessed separately in our monitoring evaluation,  
504 but that are not independently listed under the EPBCA.  
505 **Mammals:** Includes four species listed as threatened by the IUCN, but not included under the  
506 EPBCA. EPBCA listed marine species excluded.  
507

508 **Tables S2-S10. Scoring criteria for each of the nine metrics used to assess monitoring**  
 509 **quality.** Adopted from Woinarski (2018) A framework for evaluating the adequacy of  
 510 monitoring programs for threatened species, In S. Legge, D.B. Lindenmayer, N.M. Robinson,  
 511 B.C. Scheele, D.M. Southwell, B.A. Wintle editors. Monitoring Threatened Species and  
 512 Ecological Communities. CSIRO Publishing, Melbourne.

513

514 Table S2. Metric 1: Fit-for-purpose

Score	Score basis
5	sampling protocol effectively applies all relevant methods that have been demonstrated to optimise detectability and population estimation at sampled sites
4	sampling protocol is based on one or more methods that can reliably detect species
3	sampling protocol is likely to detect species if present, but will provide limited information on abundance
2	not known if sampling protocol will reliably detect species if present
1	sampling protocol is an unreliable approach to demonstrating occurrence or abundance
0	no monitoring

515

516 Table S3. Metric 2: Coverage

Score	Score basis
5	monitoring undertaken comprehensively across range
4	monitoring undertaken representatively at many sites across range
3	monitoring undertaken at several sites across range, but significant components not monitored
2	monitoring at a few sites, not necessarily representative
1	monitoring at one site only (except where this is the only site of occurrence)
0	no monitoring

517

518 Table S4. Metric 3: Periodicity

Score	Score basis
5	monitoring at least annually at major site(s)
4	monitoring at 2–3 year intervals
3	monitoring at 4–9 year intervals
2	monitoring at >10 year intervals
1	no repeat sampling
0	no monitoring

519

520

## 521 Table S5. Metric 4: Longevity

Score	Score basis
5	monitoring extending back for >30 years, and there is an assurance of ongoing commitment
4	monitoring extending back for >20 years, and there is some indication of ongoing commitment
3	monitoring extending back for >10 years
2	monitoring extending back for >5 years
1	monitoring established only in last 1–5 years, no consideration of future sampling
0	no monitoring

522

## 523 Table S6. Metric 5: Design quality

Score	Score basis
5	high statistical power to detect small (e.g. 5%) change in population size over a timeframe relevant to conservation context
4	sufficient statistical power to reliably detect moderate (e.g. 30% change) in population size
3	reasonable design but low statistical power (e.g. unlikely to reliably detect 50% change in population size)
2	rudimentary design but resulting in sufficient records to suggest broad changes in abundance
1	typically <i>ad hoc</i> with few records
0	no monitoring

524

## 525 Table S7. Metric 6: Coordination

Score	Score basis
5	monitoring activities tightly integrated across sites with unambiguous overall responsibility and consistent sampling methodologies
4	broad consistency in sampling protocols across sites; some linking of results across sites
3	some links established between monitoring projects at different sites
2	range of monitoring projects compiled, but no linking of results, and inconsistent sampling protocols across sites
1	no apparent coordination in monitoring activities, design or reporting
0	no monitoring

526

527

## 528 Table S8. Metric 7: Data availability

Score	Score basis
5	all relevant data are collated, readily available and up to date on well-established and publicly accessible sites, with robust analysis and interpretation
4	all relevant data readily available and up to date on publicly accessible sites
3	reasonably easy to find some information on monitoring results, either through websites or published reports or scientific papers
2	some information may be available, but difficult to access readily
1	monitoring information largely unobtainable by others
0	no monitoring

529

## 530 Table S9. Metric 8: Management linkage

Score	Score basis
5	monitoring closely linked to adaptive management, providing an explicit measurement of threat impacts and/or management performance; monitoring includes inbuilt triggers or review that prompt management responses
4	monitoring design explicitly assesses different threat impacts and management responses, and has some links to management agency; triggers (if existing) are weakly defined and do not necessarily provoke management response
3	monitoring programs provide some consideration of effects of different management regimes; no defined triggers
2	monitoring program may provide weak inference about management, but no clear links to adaptive management; no triggers
1	monitoring program not capable of assessing management effectiveness
0	no monitoring

531

## 532 Table S10. Metric 9: Demographic parameters

Score	Score basis
5	monitoring program includes detailed assessment of relevant life history parameters (such as reproductive success, mortality rates and their causes), which can identify weak points in biology and interventions for management action, and monitoring data allow predictive population modelling
4	monitoring includes reliable information on at least one relevant life history parameter
3	monitoring includes some consideration of at least one relevant life history parameter
2	monitoring parameters are restricted to incidence or abundance, but in a manner that may allow reasonable inference about some life history parameters
1	monitoring parameters are restricted to incidence or abundance
0	no monitoring

533

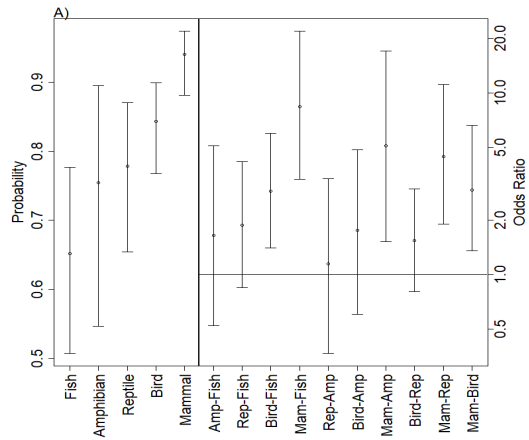
534 **Table S11.** Leave-One-Out-Cross-Validation Information Criteria for each of the eight models  
 535 considered for each of the 11 response variables (monitoring presence/absence, total score and  
 536 individual score for each of the nine metrics). Best ranked model highlighted with bold text.

	Null	Class(C)	Recovery Plan (RP)	EPBCA Status	RP + EPBCA	C + EPBCA	C + RP	CP + RP + EPBCA
Monitoring (pre/abs)	467.4	444.1	436.1	449.6	431.4	431	420.3	<b>416</b>
Total across nine metrics	2145.3	2095.4	2127.6	2135.6	2112.9	2085.2	2076.8	<b>2061.4</b>
1. Fit for purpose	916.2	897.2	892.2	915.6	885.5	897.4	874.3	<b>868.5</b>
2. Coverage	955.5	926.3	948.5	935.5	919.8	911.7	921.2	<b>899.9</b>
3. Periodicity	1022.2	953.5	1014.9	1021.2	1013	950.7	945.6	<b>943.3</b>
4. Longevity	1016.3	997	997.5	1020.1	1003.3	1002.8	<b>974.9</b>	980.2
5. Design quality	988.6	913.6	977.6	981.1	970.7	904.5	901	<b>892.5</b>
6. Coordination	1093	939.9	1092.8	1078.2	1076.8	<b>918.8</b>	939.3	918.7
7. Data availability	897.8	898.1	891.1	891.8	883.4	890.1	891.1	<b>880.9</b>
8. Management linkage	999.8	1002.6	999	994.8	<b>988</b>	1000.2	1000.7	993.3
9. Demographic parameters	1006.5	993.3	995.9	1001.9	989.8	991.7	984.1	<b>980.5</b>

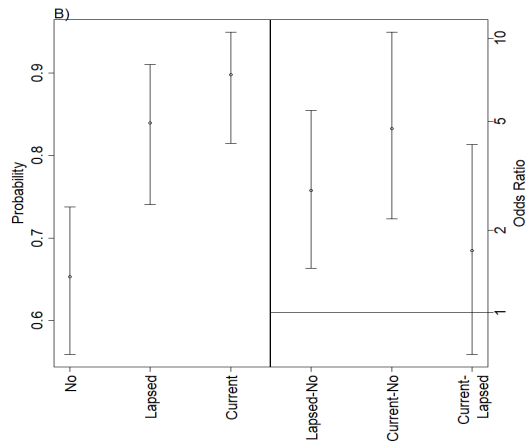
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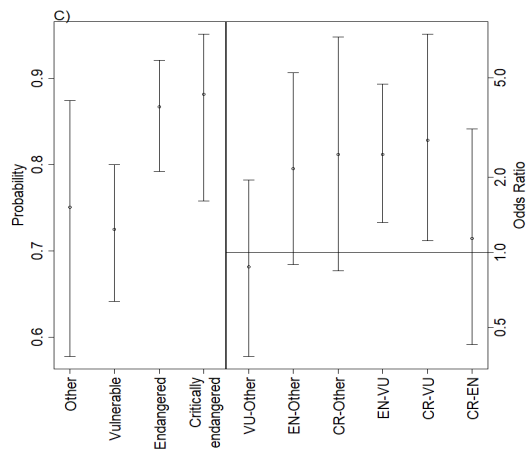
539 **Figures S1-S11. Model predictions**



540



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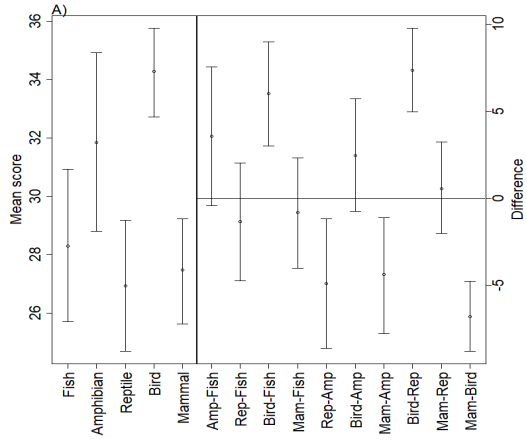
542

543 Figure S1. Probability of presence/absence of monitoring for Australian threatened taxa by (A)  
 544 taxonomic class, (B) recovery plan status, and (C) conservation status. In each case, the  
 545 probability of monitoring was predicted at average values for the other two predictors in the

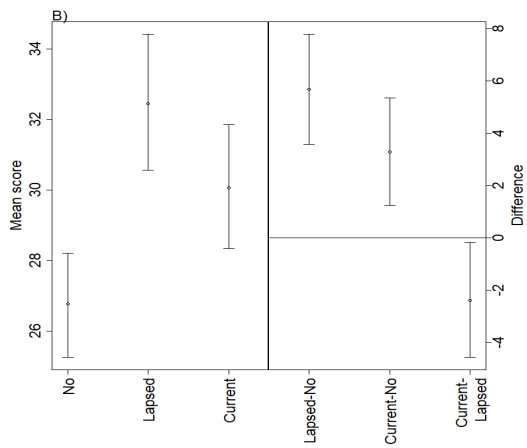
546 model. The right side of each panel depicts the odds ratio and 95% credible interval comparing  
547 each level of the factor.



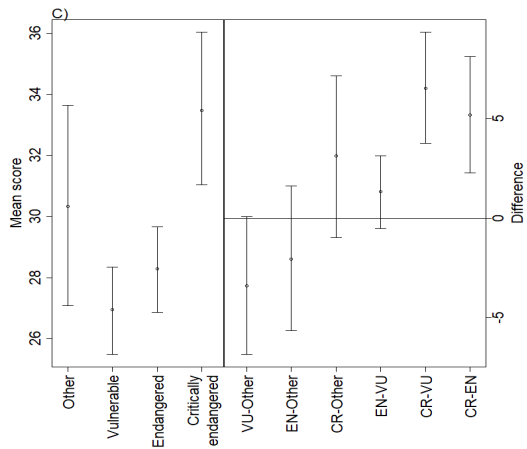
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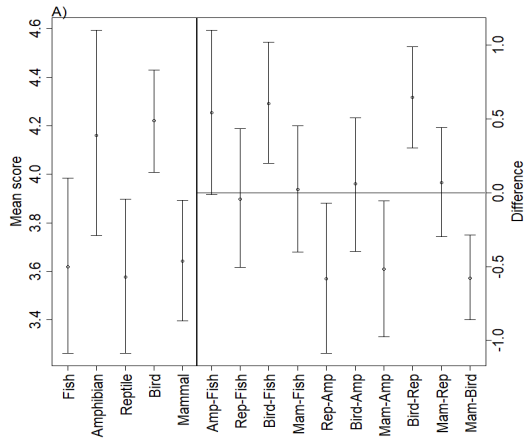


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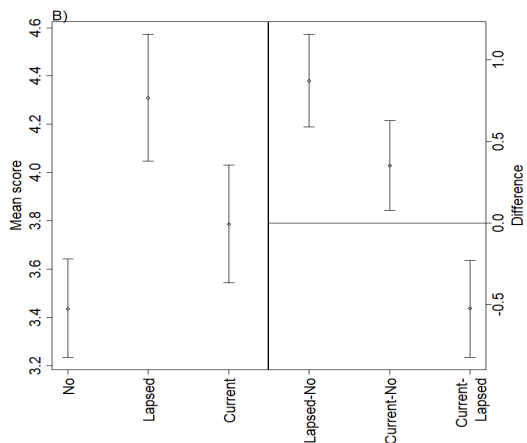


551 Figure S2. Predicted mean total score for monitoring quality for Australian threatened taxa by  
552 (A) taxonomic class, (B) recovery plan status, and (C) conservation status. In each case,  
553 predictions were made at average values for the other two predictors in the model. The right side  
554 of each panel depicts the mean difference and 95% credible interval comparing each level of the  
555 factor.

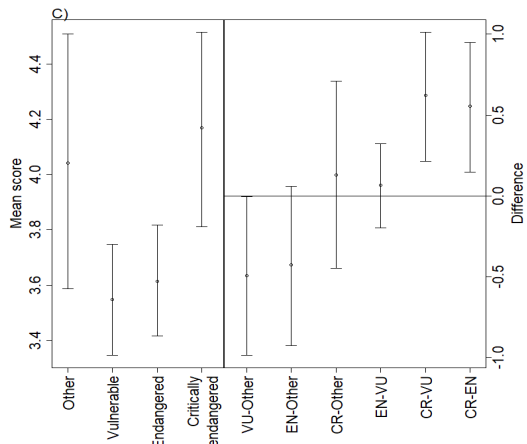
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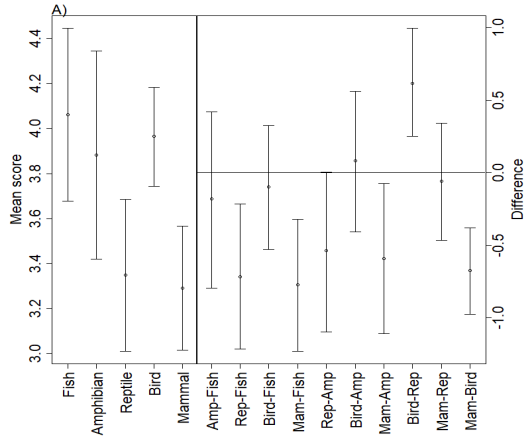


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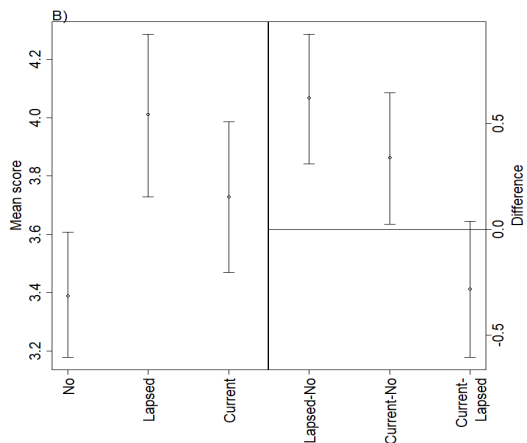


559 Figure S3. Predicted score for fit-for-purpose for monitoring programs for Australian threatened  
560 taxa by (A) taxonomic class, (B) recovery plan status, and (C) conservation status. In each case,  
561 predictions were made at average values for the other two predictors in the model. The right side  
562 of each panel depicts the mean difference and 95% credible interval comparing each level of the  
563 factor.

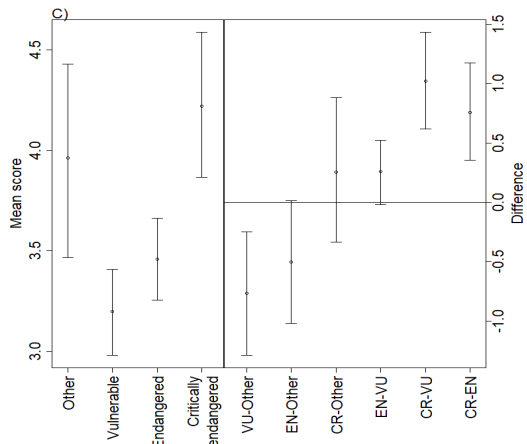
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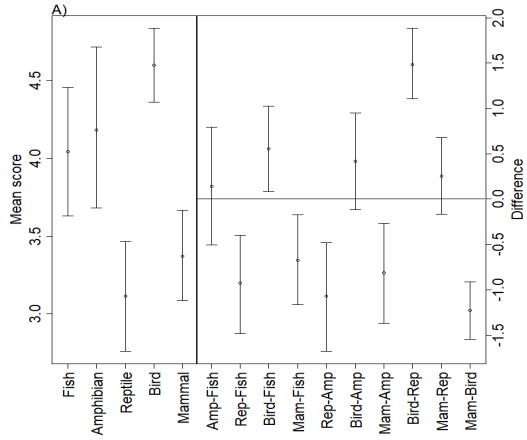


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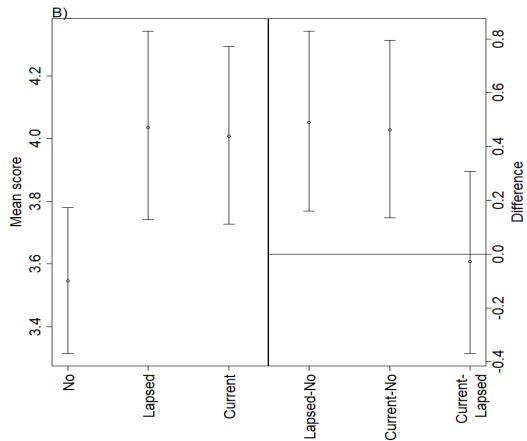


567 Figure S4. Predicted score of monitoring coverage for Australian threatened taxa by (A)  
568 taxonomic class, (B) recovery plan status, and (C) conservation status. In each case, predictions  
569 were made at average values for the other two predictors in the model. The right side of each  
570 panel depicts the mean difference and 95% credible interval comparing each level of the factor.

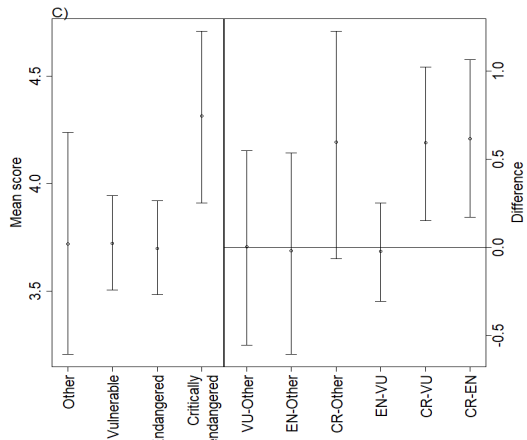
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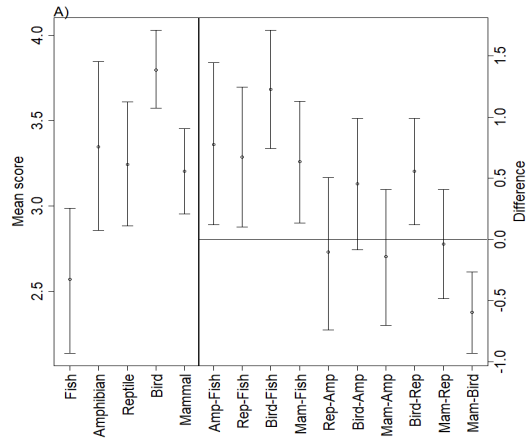
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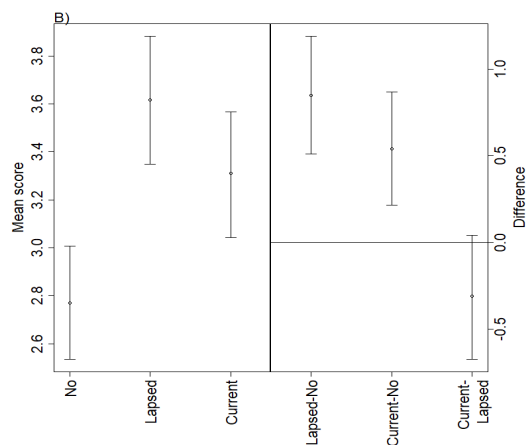
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Figure S5. Predicted score for periodicity of monitoring for Australian threatened taxa by (A) taxonomic class, and (B) recovery plan status, and (C) conservation status. In each case, predictions were made at average values for the other two predictors in the model. The right side of each panel depicts the mean difference and 95% credible interval comparing each level of the factor.

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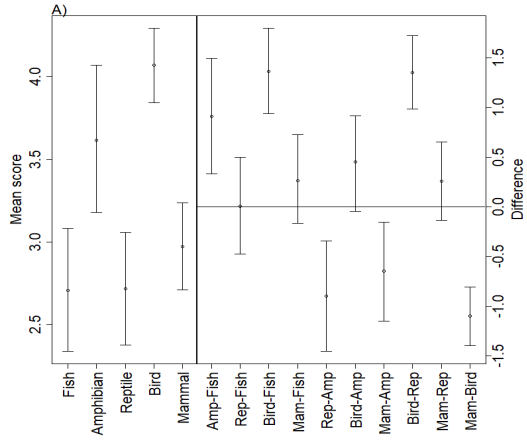
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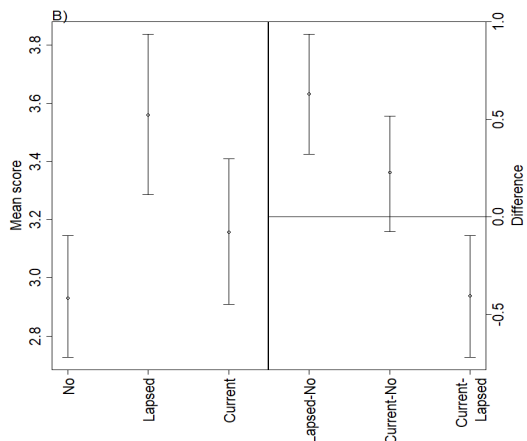
581 Figure S6. Predicted score for longevity of monitoring for Australian threatened taxa by (A)  
582 taxonomic class, and (B) recovery plan status. In each case, predictions were made at average  
583 values for the other predictor in the model. The right side of each panel depicts the mean  
584 difference and 95% credible interval comparing each level of the factor.

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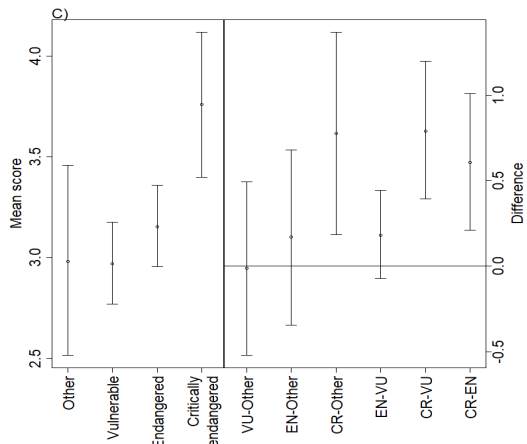
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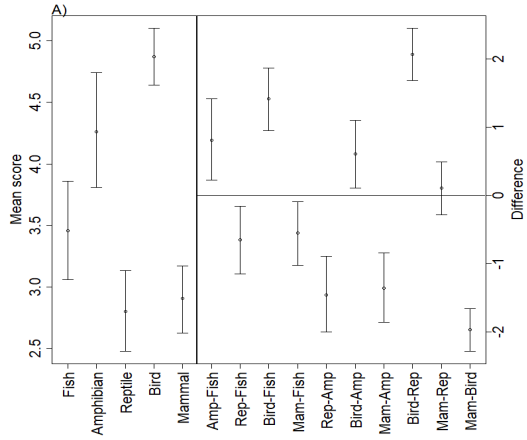


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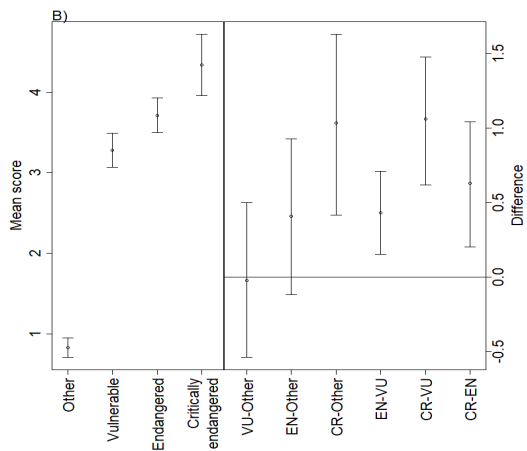


589 Figure S7. Predicted score for monitoring design quality for Australian threatened taxa by (A)  
 590 taxonomic class, (B) recovery plan status, and (C) conservation status. In each case, predictions  
 591 were made at average values for the other two predictors in the model. The right side of each  
 592 panel depicts the mean difference and 95% credible interval comparing each level of the factor.

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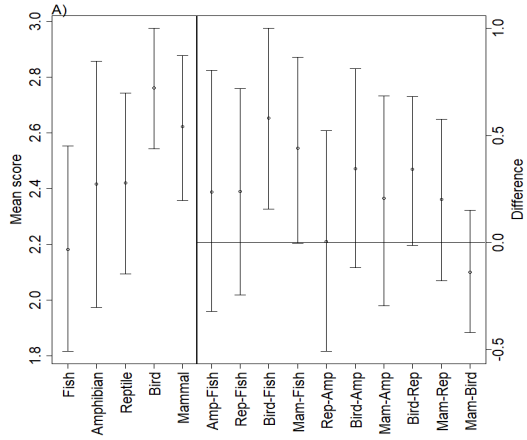


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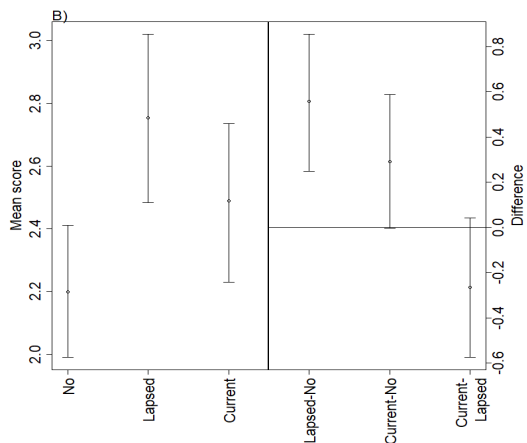
596 Figure S8. Predicted score for monitoring coordination for Australian threatened taxa by (A)  
597 taxonomic class, and (B) conservation status. In each case, predictions were made at average  
598 values for the other predictor in the model. The right side of each panel depicts the mean  
599 difference and 95% credible interval comparing each level of the factor.

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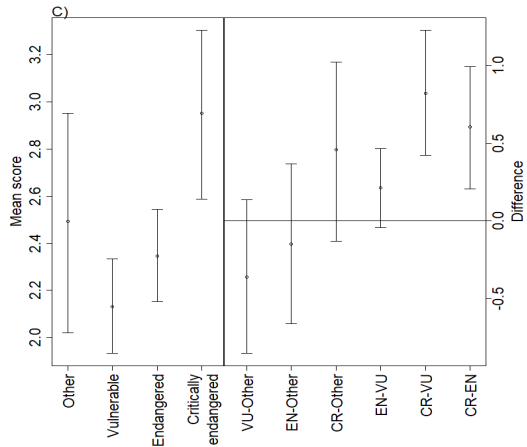
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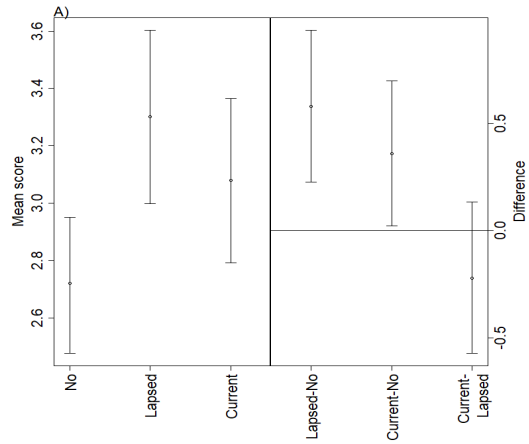


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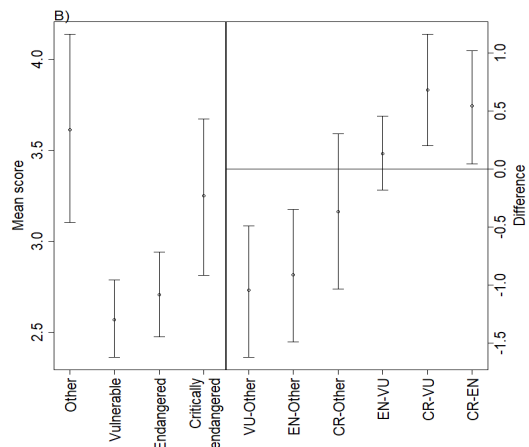


604 Figure S9. Predicted score for monitoring data availability for Australian threatened taxa by (A)  
 605 taxonomic class, (B) recovery plan status, and (C) conservation status. In each case, predictions  
 606 were made at average values for the other two predictors in the model. The right side of each  
 607 panel depicts the mean difference and 95% credible interval comparing each level of the factor.





608

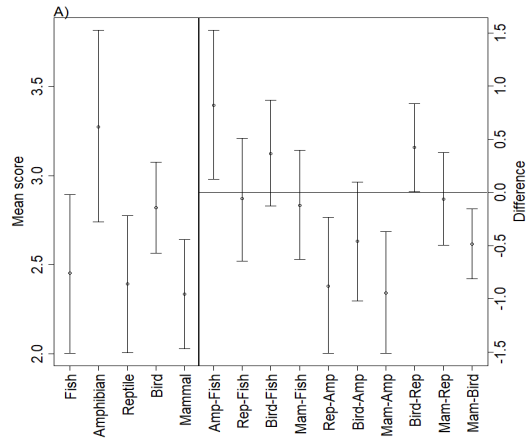


609

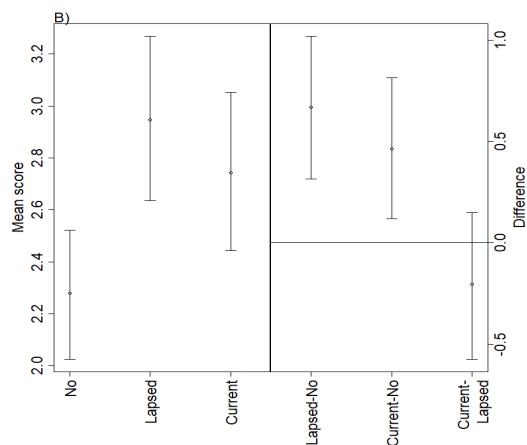
610 Figure S10. Predicted score for management linkages for Australian threatened taxa by (A)  
 611 taxonomic class, and (B) conservation status. In each case, predictions were made at average  
 612 values for the other predictor in the model. The right side of each panel depicts the mean  
 613 difference and 95% credible interval comparing each level of the factor.

614

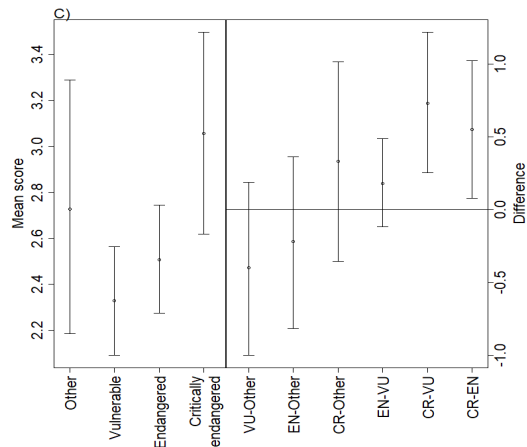
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619 Figure S11. Predicted score for consideration of demographic parameters in monitoring for  
 620 Australian threatened taxa by (A) taxonomic class, (B) recovery plan status, and (C)  
 621 conservation status. In each case, predictions were made at average values for the other two  
 622 predictors in the model. The right side of each panel depicts the mean difference and 95%  
 623 credible interval comparing each level of the factor.

624

